

1 **Title:** A blueprint for national assessments of the blue carbon capacity of kelp forests applied to
2 Canada's coastline

3

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27 JM and JKB conceived and designed the study.

28 JM acquired and collated the data with assistance from BT and input from KFD, KK, KM, MHL,
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30 JM conducted the modeling and data synthesis with significant input from DKO, KK., KFD, and
31 JKB.

32 JM wrote the manuscript with significant contributions from JKB, and revisions and input from
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34

35 **Abstract (150/150 words):** Kelp forests offer substantial carbon fixation, with the potential to
36 contribute to natural climate solutions (NCS). However, to be included in national NCS
37 inventories, governments must first quantify the kelp-derived carbon stocks and fluxes leading
38 to carbon sequestration. Here, we present a blueprint for assessing the national carbon
39 sequestration capacity of kelp forests in which data synthesis and Bayesian hierarchical
40 modelling enable estimates of kelp forest carbon production, storage, and export capacity from
41 limited data. Applying this blueprint to Canada's extensive coastline, we find kelp forests store
42 an estimated 1.4 Tg C in short-term biomass and produce 3.1 Tg C yr⁻¹ with modest carbon
43 fluxes to the deep ocean. Arctic kelps had the highest carbon stocks and production capacity,
44 while Pacific kelps had greater carbon fluxes overall due to their higher productivity and export
45 rates. Our transparent, reproducible blueprint represents an important step towards accurate
46 carbon accounting for kelp forests.

47
48 **Keywords:** macroalgal forests, kelp beds, productivity, nature-based solutions, ocean climate
49 solutions, Bayesian hierarchical modeling

50

51 **Main Text (4500/4500 words):**

52

53 **Introduction**

54 As the urgency of addressing climate change intensifies, natural climate solutions (NCS)
55 involving habitat interventions to enhance natural carbon sinks have emerged as distinct
56 components of countries' mitigation strategies^{1,2}. However, most NCS assessments focus on
57 forests, grasslands, and wetlands, with less attention on the vast carbon reservoirs found in the
58 ocean^{1,3,4}. In the coastal zone, blue carbon ecosystems (BCEs)—seagrass meadows, salt marshes,
59 and mangrove forests—contribute to carbon sequestration in the ocean by converting dissolved
60 carbon dioxide (CO₂) that has been removed from the atmosphere into biomass, and by
61 promoting the burial of organic material in benthic sediments^{2,5–8}. BCE standing biomass can
62 persist for decades, and sedimentary carbon stocks can be preserved for centuries to millennia
63 when undisturbed^{9–11}. As a result, these systems remove carbon from shallow waters where it
64 would have otherwise exchanged as atmospheric CO₂ and exacerbated climate change⁸. Since

65 many BCEs have declined significantly over the past century¹², conservation and improved
66 management of these ecosystems are increasingly seen as low regret strategies for avoiding
67 further CO₂ emissions. Similarly, restoration and expansion of BCEs has also been proposed as a
68 potential strategy to enhance natural carbon sequestration in the ocean^{1,2,13}.

69
70 Kelp forests, composed of large brown seaweeds primarily from the order Laminariales, have
71 traditionally not been considered blue carbon ecosystems, due to their lack of roots and local
72 carbon burial in sediments^{14,15}. However, recent work identifies kelp forests as emerging BCEs¹⁶
73 because of their ability to efficiently assimilate CO₂¹⁷, their near global distributions^{18,19}, their
74 role as allochthonous producers of carbon-rich material, and their potential export to
75 depositional environments where sequestration occurs^{20–23}. Much like terrestrial forests, kelps
76 form expansive and highly productive vegetated canopies, with some species extending from
77 the benthos to the surface (i.e., surface kelps) and others forming dense submerged beds on
78 the seafloor (i.e., subsurface kelps). While most kelp production enters marine food webs as
79 particulate and dissolved organic carbon (POC and DOC, respectively) and is remineralized in
80 the short-term²⁴, a portion has the potential to become sequestered and stored for geological
81 timescales (i.e., 100s to 1000s of years) in various natural ocean carbon sinks^{14,17,25}. There are
82 three main pathways for kelp carbon sequestration: 1) some portion of kelp DOC is or becomes
83 refractory DOC (i.e., inaccessible to microbial communities) with residence times ranging from
84 decades to millennia when exported below the photic zone^{20,26}; 2) kelp POC in the form of
85 dislodged or fragmented kelp fronds is transported and buried in shelf sediments and the
86 sediments of other BCEs (e.g., seagrass meadows) for similar timescales^{27,28}; and 3) kelp POC

87 and DOC reaches the deep ocean (depths >200 m), where if buried can be preserved for
88 centuries to millennia because of the limited potential for resuspension to the surface
89 ocean^{20,29}.

90
91 Global assessments show considerable potential for carbon assimilation through kelp
92 productivity^{17,18}. Yet whether kelp forests can provide viable NCS remains unclear due to the
93 data gaps, process uncertainties, and the challenges associated with measuring kelp carbon
94 sequestration at relevant scales for management (e.g., regional or national)³⁰. Substantial
95 stretches of temperate and sub-arctic coastlines are suitable habitats for kelp forests^{18,19}, but
96 the actual extent of kelp forests is not fully mapped in most countries³¹ and is likely to exhibit
97 seasonal and interannual variability in both extent and productivity³². Kelp-derived carbon
98 stocks and fluxes (i.e., biomass, productivity, export, and sedimentary accumulation rates)
99 leading to carbon sequestration are also uncertain because of natural variation and incomplete
100 knowledge of their distribution, production, and POC and DOC fates. Moreover, not all
101 exported kelp POC will be stored for long enough to be considered relevant for climate change
102 mitigation (i.e., >100 years) and not all kelp POC that is stored will fall into existing carbon
103 accounting and verification standards (i.e., within verifiable and governable reservoirs inside a
104 country's exclusive economic zone; EEZ)^{14,15}. Given these uncertainties, new approaches are
105 needed to estimate the current carbon sequestration capacity of kelp forests at national scales.

106
107 To facilitate accurate carbon accounting, we present a blueprint for producing national
108 assessments of the blue carbon capacity of kelp forests (Fig. 1). Combining kelp data collation

109 with Bayesian hierarchical modeling, this transparent and reproducible analytical framework
110 estimates the carbon sequestration capacity of kelp forest ecosystems while explicitly
111 acknowledging the inherent data limitations and uncertainties that most countries face in this
112 regard. We apply this blueprint to Canada—a country accounting for 16.2% of the worlds
113 coastline³³—with expansive kelp forest ecosystems in the Atlantic, Pacific, and Arctic oceans.
114 Two major surface canopy species, giant (*Macrocystis pyrifera*) and bull (*Nereocystis luetkeana*)
115 kelp, form extensive floating forests along the Canadian Pacific, while subsurface species from
116 the genera *Laminaria*, *Saccharina*, *Alaria*, *Agarum*, and others form dense submerged beds on
117 their own, or as an understory below surface kelps, along substantial stretches of the Canadian
118 coastline⁴⁰. Our study enables the inclusion of kelp forest ecosystems into national NCS
119 inventories in Canada and other countries with these important coastal ecosystems.

120

121 **Results**

122 Kelp forest blue carbon blueprint

123 Our blueprint for national assessments of the blue carbon capacity of kelp forests involves: 1)
124 compiling and synthesizing available kelp data and identifying data gaps, 2) evaluating the
125 potential for natural variation in the carbon stocks and carbon production rates for kelp
126 species, 3) developing initial estimates of the standing carbon stock, production, and export
127 capacity of kelp forests to deep ocean sinks, and 4) refining assessments based on new
128 information and data (Fig. 1). For reproducibility, we provide a blueprint workflow and
129 methodology for conducting an extensive collation of available datasets on the areal extent,
130 canopy biomass, and NPP of kelp forests (Appendix A). We also provide R scripts that enable

131 users to estimate the posterior mean carbon stocks and production rates of different kelp
132 species based on limited available data and prior information using Bayesian hierarchical
133 models ('Brms' package), as well as templates for scaling up per-area estimates to a national
134 scale (Appendix B). Below we illustrate the blueprint's utility through an application to Canada.

135

136 First blueprint application: Canadian kelp forests

137 Data collation

138 We first compiled a database of kelp records from 36 published studies and monitoring
139 programs (Appendix C: Table C1; Fig. C1) describing the areal extent, abundance (i.e., biomass
140 and density), and NPP of subtidal kelp forest species across Canada's Pacific, Atlantic and Arctic
141 coasts (Fig. 1, Step 1). Our search targeted available data for surface kelp species found on the
142 Pacific coast and subsurface kelps found across Canada's three coasts, revealing that eleven of
143 the 18 subtidal kelp species in Canada had sufficient data records to be included in further
144 analyses. These include the two surface kelp species and seven of the 15 subsurface kelps
145 located on the Pacific coast, five of the seven subsurface species found on the Arctic coast, and
146 three of the five species found on the Atlantic coast (Table C2).

147

148 Kelp forest extents

149 *Subsurface kelps:* Next, since synoptic maps were unavailable, we produced high, mid, and low
150 estimates of the potential extent of subsurface kelp forests in Canada using available depth,
151 substrate, and kelp percent cover data (Table C3). To determine a hypothetical maximum limit
152 for where subsurface kelp forests could occur in Canada, we calculated the area of rocky reefs

153 (i.e., bedrock and boulders habitats) from mean-low-low-water out to 20 m water depth. With
154 only this depth and substrate constraint, we found that subsurface kelp forests could cover up
155 to 6.3 million hectares (Mha) (Table 1). Most of the kelp forest distribution (approximately 71%)
156 was estimated to occur in the Arctic (5.5 Mha), while Atlantic and Pacific kelp forests covered
157 1.3 and 0.5 Mha, respectively (Appendix C: Table C3). Given that kelp do not always completely
158 cover benthos, we then produced more constrained estimates for subsurface kelps, using
159 available field surveys of kelp percent cover (Table C1), acknowledging that kelp percent cover
160 can also vary annually and seasonally. We determined an upper biologically constrained extent
161 of subsurface kelps by multiplying the maximum potential extent (described above) and the
162 upper quartile of observed kelp percent cover at peak canopy biomass (May – August) across
163 sites and years on each coast. We also determined a lower biologically constrained extent of
164 subsurface kelp forests by multiplying the maximum potential extent by the lower quartile
165 percent cover values, across years and sites, on each coast. Although the exact extent of
166 subsurface kelp forests is still unknown, we estimated that the true extent falls between 0.8
167 and 3.9 Mha (Table C3).

168

169 *Surface kelps:* As a particular case found on the Pacific Coast, we also produced high, mid, and
170 low estimates specifically for surface kelp forests using available remote sensing and aerial
171 surveys (Fig. C2). For the high estimate, we calculated the area of rocky reefs from mean-low-
172 water out to 10 m water depth (Fig. C3). As an upper bound estimate, we used historical
173 shoreline maps derived from oblique aerial survey imagery conducted by the British Columbia
174 Shore Zone Survey from 2004-2007 to identify shallow rocky reefs that were previously covered

175 by surface kelp forests. Finally, we used recent global surface canopy maps derived from
176 Sentinel-2 satellite imagery from 2015 to 2019 as a low bound estimate. According to this
177 analysis, surface kelp forests on the Pacific coast of Canada could cover up to 0.3 Mha, but
178 more conservatively cover between 0.005 and 0.11 Mha (Table C3).

179

180 Per-area carbon stocks and production rates of kelp species

181 Bayesian hierarchical models revealed significant differences in per area carbon stocks and
182 productivity within and among kelp species in Canada (Fig. 1, Step 2). On average, surface kelps
183 tended to have higher values than subsurface species (Fig. 2). Giant and bull kelp, stored more
184 carbon per area in their canopy biomass than six of the seven subsurface species (1.30 Mg C ha⁻¹
185 and 0.95 Mg C ha⁻¹, respectively), with over 80% conditional support for differences amongst
186 the posterior mean predictions (Fig. 2a, Appendix C: Table C4). Giant and bull kelp also had the
187 highest annual carbon production rates per area (7.26 Mg C ha⁻¹ yr⁻¹ and 6.35 Mg C ha⁻¹ yr⁻¹,
188 respectively), producing more than twice the amount of carbon per year of other kelp species
189 (Fig 2b; Table C4). While certain subsurface kelps (e.g., *Saccharina latissima*) had comparable
190 estimated carbon stocks and production rates to surface kelps, most had much lower estimated
191 carbon stocks per-area—ranging from 0.01 Mg C ha⁻¹ (*Pleurophycus gardneri*) to 0.66 Mg C ha⁻¹
192 (*Pterygophora californica*)—and carbon production rates per-area —ranging from 0.08 Mg C ha⁻¹
193 yr⁻¹ (*Agarum clathratum* / *Neoagarum fimbriatum*) to 3.18 Mg C ha⁻¹ yr⁻¹ (*Laminaria digitata* /
194 *Hedophyllum nigripes*) (Fig. 2).

195

196 Per-area carbon stocks, production, and export rates of kelp forests by coast

197 Across Canada's three coasts, we found considerable variation in the estimated per-area carbon
198 stock and production rates of kelp forests due to differences in species composition and peak
199 biomass (Fig. 3). Overall, Pacific kelp forests had the largest estimated carbon stocks per-area
200 (1.2 Mg C ha^{-1}), along with the largest number of kelp species ($N=17$), and the highest
201 estimated annual carbon production rates ($6.7 \text{ Mg C ha}^{-1} \text{ yr}^{-1}$) (Fig. 3a). In comparison, Atlantic
202 and Arctic kelp forests had lower kelp diversity ($N=7$ and 5 , respectively) and a lower estimated
203 carbon stock potential (0.4 and 0.8 Mg C ha^{-1} , respectively), as well as much lower annual
204 carbon production rates ($2.7 \text{ Mg C ha}^{-1} \text{ yr}^{-1}$ and $1.3 \text{ Mg C ha}^{-1} \text{ yr}^{-1}$, respectively) (Fig. 3b).

205
206 As an approximation of the upper limit for carbon sequestration occurring in the deep ocean
207 from Canada's kelp forests, we also estimated the per-area rate of detrital export from kelp
208 forests to beyond the continental shelf break (i.e., the 200-m isobath) according to global ocean
209 transport estimates. Approximately 22.0% (SD = 12.0%) of kelp detritus is likely to reach the
210 continent shelf break before decomposing in the Pacific coast compared to 10.8% (SD= 6.7%) in
211 the Atlantic and 8.8% (SD= 2.8%) in the Arctic (Table B7). This implies that approximately 1.5
212 $\text{Mg C ha}^{-1} \text{ yr}^{-1}$ ($0.7 - 2.3 \text{ Mg C ha}^{-1} \text{ yr}^{-1}$) could be exported from Pacific kelp forests compared to
213 $0.3 \text{ Mg C ha}^{-1} \text{ yr}^{-1}$ ($0.2 - 0.4 \text{ Mg C ha}^{-1} \text{ yr}^{-1}$) from Atlantic kelp forests and $0.1 \text{ Mg C ha}^{-1} \text{ yr}^{-1}$ (0.01
214 $- 0.2 \text{ Mg C ha}^{-1} \text{ yr}^{-1}$) from Arctic kelp forests (Fig. 3c).

215

216 *First national estimates for Canada's kelp forests*

217 Finally, to produce national estimates of the carbon sequestration capacity associated with
218 Canada's kelp forests, we combined the kelp forest extent estimates with the per-area carbon

219 stock, carbon production, and carbon export estimates on each coast (Fig. 1, Step 3). For a
220 conservative scenario, assuming kelp forests are at their median areal extent, we estimate that
221 Canadian kelp forests have a standing carbon stock capacity of 1.4 Tg C (0.6 - 2.8 Tg C) and an
222 annual carbon production capacity of 3.1 Tg C yr⁻¹ (1.1 – 6.3 Tg C yr⁻¹), approximately 0.2 Tg C
223 Yr⁻¹ (0.04 – 0.4 Tg C yr⁻¹) of which could be transported to and sequestered in the deep ocean
224 (Fig. 4). However, in the most optimistic scenario, where kelp forests are at their maximum
225 potential extent, these figures increase to a national standing stock capacity of 4.4 Tg C, an
226 annual carbon production capacity of 11.6 Tg C yr⁻¹, and an annual carbon export capacity of 1.0
227 Tg C yr⁻¹ to the deep ocean. Arctic kelp forests had the greatest overall carbon stock (1.3 Tg C;
228 0.6 – 3.5 Tg C) and production capacity (2.1 Tg C yr⁻¹; 1.0 – 5.8 Tg C yr⁻¹) (Fig. 4a - b) because of
229 their disproportionately larger areal extents. However, kelp forests in the Pacific had the
230 highest estimated capacity for carbon sequestration via export to the deep ocean (0.15 Tg C yr⁻¹
231 ¹, 0.01 – 0.5 Tg C yr⁻¹) because of the higher per-area carbon production rates and potential for
232 detrital transport beyond the shelf break (Fig. 4c).

233

234 **Discussion**

235 National assessments of BCEs, such as seagrasses, salt marshes, and mangroves, are becoming
236 more prevalent^{2,34}, paving the way for their incorporation into NCS inventories. However,
237 comparable evaluations for kelp forests are currently unavailable for nearly 90% of the 150
238 countries with kelp forests¹⁴, due to existing data gaps and the difficulty of accurately
239 estimating the kelp-derived carbon stocks and fluxes leading to sequestration in various ocean
240 sinks (i.e., DOC pools, shelf sediments, and the deep ocean). Our reproducible blueprint,

241 applied to Canadian kelp forests, has important implications for other countries looking to
242 account for kelp forests as NCS.

243

244 ***Potential for Canada's kelp forests as an NCS***

245 Our assessment found the carbon production capacity of Canadian kelp forests to be
246 substantial (3.1 Tg C yr⁻¹; 1.1 - 11.5 Tg C yr⁻¹). This figure is low compared to recent global
247 estimates of kelp carbon production (~1.5% of global estimated NPP) but these are not directly
248 comparable as we used a more conservative depth cut-off when calculating the extent of kelp
249 forests (20m compared to 30m)¹⁸. Additionally, we found that Canadian kelp forests provide a
250 clear pathway for sequestering and storing carbon in the deep ocean (0.2 Tg C yr⁻¹, 0.04 to 1.0
251 Tg C yr⁻¹). Realised carbon sequestration of kelp-derived carbon could be even greater when
252 accounting for kelp carbon entering refractory DOC pools in the deep ocean. Compared to
253 terrestrial ecosystems, kelp forests are likely to play a more modest role in the overall NCS
254 components of Canada's climate change mitigation strategy³. As examples, conservation
255 pathways for grasslands, peatlands, and forests have been estimated to sequester 3.5 Tg C Yr⁻¹,
256 2.8 Tg C yr⁻¹, and 2.2 Tg C yr⁻¹, respectively³. Nevertheless, we found that kelp forests could
257 have comparable carbon sequestration benefits to freshwater mineral wetlands (0.8 Tg C yr⁻¹)³
258 and other BCES, such as eelgrass meadows (0.2 Tg C yr⁻¹), and tidal marshes (0.8 Tg C yr⁻¹)
259 (Table 1), thus warranting further consideration in Canada's NCS inventories.

260

261 We found contrasting patterns of carbon production, storage, and export across Canada's
262 coastlines, reflecting different kelp species assemblages and environmental conditions across

263 these vast areas. Notably, the per-area carbon production capacity of Pacific kelp forests
264 exceeded the global averages for subtidal kelps, intertidal seaweeds, subtidal red seaweeds,
265 and floating seaweeds (e.g., *Sargassum* spp)¹⁸. While the Arctic had the highest total kelp
266 standing carbon stock and production capacity, due to their extensive coastline and wide
267 continental shelf, Pacific and Atlantic coasts had a higher capacity for kelp carbon sequestration
268 in the deep ocean overall, because of their higher per-area rates of kelp carbon production and
269 hydrological export. However, kelp forests on all three coastlines showed some potential to
270 sequester carbon in the deep ocean, highlighting their potential role in Canada's NCS
271 inventories.

272

273 Our assessment revealed considerable data gaps across all elements of our analysis,
274 underscoring the need for kelp monitoring programs at a national scale. Given the lack of
275 comprehensive habitat maps, we needed to make assumptions about the current extent of kelp
276 forests, including assumptions about the maximum depth limit of kelp forests, the prevalence
277 of rocky reefs, and the kelp occupancy and abundance across Canada's coastline. We also could
278 not account for ecological driver (e.g., urchins) that likely limit kelp extents in certain areas³⁵.

279 When estimating per-area carbon stocks and production rates, we faced significant data
280 limitations for many kelp species, especially the subsurface kelps, leading to large credible
281 confidence intervals for many species. Additionally, we needed to make assumptions about the
282 relative abundance of species in kelp forests when extrapolating standing carbon stock,
283 production, and export estimates to the coast-wide scale. Lastly, given the complete lack of
284 data on the accumulation of kelp derived carbon in shelf and deep ocean sinks, we relied on

285 hydrological export estimates from ocean transport models to approximate carbon
286 sequestration rates beyond the continental shelf breaks (as only one potential pathway of kelp
287 carbon sequestration). A sensitivity analysis revealed that the maximum depth limit and the
288 hydrological export rates are likely to have the strongest influence on national estimates (Fig.
289 C4 – C6), suggesting these datasets should be the highest priority for future research. However,
290 greater investment in collecting of kelp species abundance, composition, and net primary
291 productivity data is also needed to estimate kelp carbon sequestration more accurately
292 nationally. Notably, many of these data can be and are already being collected by coastal
293 communities and First Nations in Canada (e.g., through the Marine Plan Partnership program)³⁶,
294 creating future opportunities for collaborative research efforts.

295

296 ***General considerations of kelp NCS***

297 While highlighting kelp forests in Canada as a potential asset for storing and sequestering
298 atmospheric carbon, our study has broader implications for developing NCS in other data-
299 limited countries with kelp forests. First, our findings underscore the importance of elucidating
300 and considering the full pathways of carbon production, export, and storage. For example,
301 despite the seemingly higher total carbon stocks and production capacity of Canada’s Arctic
302 kelp forests, a comprehensive understanding of the area-based production and export potential
303 of distinct coastlines was needed to reveal the greater contributions of Pacific and Atlantic kelp
304 forests to carbon export fluxes to the deep ocean. However, it is possible that Arctic kelp
305 forests could play a more important role in shelf carbon sequestration, particularly given the
306 Arctic’s expansive shallow continental shelf and the greater potential for preservation due to

307 cold temperatures³⁷. Additionally, our findings emphasize the potential importance of spatial
308 and temporal variation in kelp carbon cycling. In Canada, potential export rates varied by an
309 order of magnitude difference(0.9 to 33.6%)depending on the ecoregion³⁸. Additionally, kelp
310 species showed considerable variation in their estimated carbon stocks and production rates,
311 which may increase further as more spatially and temporally resolved data becomes available.
312 Overlooking these variables could lead to biased estimates, potentially undermining the
313 effectiveness of NCS.

314

315 Our findings imply that continued environmental changes could have varying consequences for
316 kelp carbon sequestration. Kelp degradation and deforestation have occurred globally due to
317 various anthropogenic stressors and disturbances, including overfishing, eutrophication,
318 climate change and species invasions^{39–43}. For instance, along Canada’s Pacific and Atlantic
319 coast, kelp declines have been documented after unchecked urchin grazing and intensifying
320 marine heatwaves^{35,44}, while many kelp forests in Atlantic Canada have transitioned to beds
321 dominated by algal turfs due to the combined effects of warming temperatures and
322 interactions with invasive species^{40,45,46}. These changes are likely to have severe consequences
323 for associated biodiversity and other ecosystem functions (e.g., fisheries production), and they
324 may also disproportionately reduce the capacity of kelp forests to produce and export carbon.

325

326 Variation in the response of different kelp species to climate change may also have important
327 implications for understanding the impacts of kelp species redistribution on carbon
328 sequestration⁴⁷. As kelp distributions are altered by warming temperatures, there could be

329 considerable changes in kelp community composition^{48,49}; additionally, more frequent marine
330 heatwaves may lead to local extirpations of kelp species, which could impact carbon production
331 and storage patterns^{32,50}. It is possible these changes could lead to enhanced kelp carbon
332 sequestration at the cold edge of species' ranges. For instance, models from the Arctic show
333 the possibility of range expansions for *S. latissima*, *A. clathratum*, and *A. esculenta* with the loss
334 of sea ice and warming ocean temperatures⁵¹, which could further increase overall carbon
335 production in this region. However, warming ocean temperatures could also lead to faster
336 decomposition rates⁵², and it is unknown whether these gains in suitable area and productivity
337 would offset losses occurring at the warmer range edges⁵³ or locally warm hotspots^{32,40,54}.
338 Ultimately, expanded monitoring datasets and better forecasting models are needed to
339 understand the full scope of climate impacts on kelp carbon sequestration.

340

341 ***Applying the blueprint***

342 Our blueprint can be applied to other countries, providing a roadmap for evaluating carbon
343 stock, production, and export capacity in other kelp-dominated systems. Developed in a
344 country with an with extensive coastlines, diverse kelp communities, and complex
345 oceanographic settings, our approach is useful for evaluating kelp forests wherever there is
346 data on the areal extent, abundance, and net primary productivity of kelp species (see Fig. 1).
347 For coastal countries where comprehensive maps of kelp forest extent are not yet available,
348 coarse approximations could be obtained using global data on coastal bathymetry and existing
349 global species distribution models^{18,50}. Additionally, publicly available data on the NPP of kelps
350 can be extrapolated from other systems and global models, and used as prior information and

351 data in the absence of regional datasets⁵⁵. In the absence of regional models, empirical
352 measurements of rates of kelp carbon export can be supplemented with ocean transport
353 models, which can help approximate coastal to open ocean transport to various long-term sinks
354 (e.g. ^{23,56}). However, current export models will require a thorough interrogation with *in-situ*
355 experiments and observation studies.

356

357 Through integrating Bayesian hierarchical models with extensive data collation and synthesis,
358 our approach addresses prevailing challenges associated with estimating species-specific
359 carbon stocks and productivity rates, including data scarcity and the unknown potential for
360 variability. One key advantage lies in the ability of Bayesian models to leverage prior
361 information about the known range and variability of kelp productivity from related species and
362 systems when making posterior predictions. Bayesian hierarchical models can also allow for
363 incorporating different forms of measurement error (e.g., standard deviations in field
364 measurements across years) for a more transparent accounting of the residual uncertainty.
365 Furthermore, our approach produces national scale estimates in terms of a conservative range
366 from a lower bound to a maximum potential as determined by prior information and data. By
367 providing this range, our approach acknowledges the inherent complexities and variability of
368 kelp ecosystem dynamics by providing a range estimate, offering decision-makers a more
369 comprehensive and yet cautious perspective to guide policy and management strategies.

370

371 ***Priority Research Directions***

372 Application of our blueprint to Canada's kelp forests underscores general uncertainties and
373 priority research directions for countries attempting to incorporate kelp forests in their NCS
374 inventories. First, our assessment estimates the standing carbon production capacity of kelp
375 forests, from which the amount of kelp POC predicted to reach the continental shelf break each
376 year can be used as a proxy for carbon fluxes to the deep ocean. However, the magnitude of
377 kelp primary production that will be exported, sequestered, and stored in the deep ocean will
378 depend on relative rates of degradation, vertical exchange, sediment accumulation, sediment
379 remineralization, and other factors^{20,29,37,57}. Additionally, the export of kelp POC to the deep
380 ocean is just one potential pathway of carbon sequestration (e.g., POC fluxes to other BCEs and
381 DOC fluxes to the deep ocean). Acquiring more comprehensive datasets on carbon
382 accumulation, export, and retention rates across various reservoirs is crucial for refining
383 estimates of the carbon sequestration benefits from these important ecosystems.

384

385 Second, significant questions remain about whether carbon sequestration by kelp forests could
386 lead to meaningful climate change mitigation benefits either through avoided emissions or
387 restoration pathways^{14,25}. Demonstrating the impact of kelp forests on air-sea CO₂ fluxes is
388 challenging and whether air-sea fluxes can offset local respiration rates in kelp forests remains
389 uncertain^{25,58}. Additionally, for kelp forests to meaningfully contribute to climate change
390 mitigation, interventions must modify GHG emissions and/or removals beyond what would
391 happen naturally (i.e., additionality). Proposed interventions, including protection of
392 threatened kelp forests, restoration of lost kelp forests, and artificial expansions of kelp forests
393 beyond their historical extents via seaweed mariculture³⁰ would need to create durable GHG

394 gains (i.e., >100 years) and fall within a country's jurisdictional boundary to be eligible for policy
395 and management action¹⁵. Comparing estimates of maximum potential carbon sequestration
396 with projections based on actual kelp distribution data could provide valuable insights into
397 opportunities for further enhancement. However, further research is needed to understand the
398 full scope of options for kelp based NCS.

399

400 Finally, as we move towards a future characterized by ocean warming and intensifying marine
401 heatwaves, there is a pressing need to understand how these changes could impact the carbon
402 sequestration capacity of kelp forests. Improved forecasting models and expanded monitoring
403 efforts are essential to anticipate changes in kelp carbon sequestration and to develop climate-
404 smart management. Integrating kelp forests into national and global climate change mitigation
405 strategies also requires robust and standardized methodologies for quantifying and verifying
406 carbon stocks and fluxes to ocean carbon sinks under future scenarios⁵⁹.

407

408 **Conclusions**

409 As nations strive to meet their net zero targets in carbon emission, incorporating kelp forests
410 into NCS inventories represents an important next step in harnessing the full potential of BCEs.
411 To that end, countries must be able to reliably estimate and predict changes in kelp carbon
412 sequestration resulting from proposed management, conservation, and restoration actions. Our
413 analytical framework offers a blueprint for developing initial assessments across a range of
414 systems, representing a significant step forward for blue carbon accounting for kelp forests.
415 Applying this framework to Canada, we show that estimated annual fluxes of kelp-derived

416 carbon to the deep ocean are at least comparable to the carbon sequestration capacity of other
417 BCEs in Canada, suggesting that kelp forests merit further consideration within Canada's GHG
418 inventories. Additionally, our study highlights many of the important considerations, data
419 needs, and uncertainties surrounding the accounting of kelp carbon sequestration at national
420 scales. Our study can serve as an important resource for policymakers, researchers, and
421 stakeholders aiming to integrate kelp forests into their climate action strategies.

422 **Methods**

423 Study area

424 Our study area spans the Pacific, Arctic, and Atlantic coasts of Canada, from mean sea level out
425 to the 20-meter depth contour. In the Pacific, this includes 25,000 km of coastline from 48 to
426 55° N; in the Arctic, 162,000 km from 51 to 83° N; and in the Atlantic, 42,000 km from 43 to 60°
427 N. These coasts support a diversity of kelp forest-forming species. Indeed, the Northeast Pacific
428 Ocean is considered the evolutionary center of origin for kelps⁶⁰ and includes >30 kelp species
429 that vary in morphology and ecological niche⁶¹. The Arctic and North Atlantic oceans were
430 subsequently colonized and recolonized following glaciation events and are now home to >10
431 kelp species^{62,63}.

432

433 Data scoping

434 We collated information and datasets from a variety of published and unpublished sources on
435 the areal extent, biomass, plant density, canopy cover, and NPP of the most common kelp
436 forest species (Table C1). We limited our search to all surface and subsurface kelp species found
437 in the subtidal zone of at least one Canadian coast, according to global species-occurrence
438 databases^{19,64}. To collate sources from the published literature, we used an existing database of
439 macroalgal NPP measurements compiled from a combination of reports, peer-reviewed studies,
440 and PhD and masters theses published between 1967 and 2021^{55,65}. Following similar search
441 criteria, we expanded this NPP database to include additional papers published between 2021 –
442 2023 for Canadian kelp species. We then used this updated NPP database to compile published
443 measurements of kelp biomass from the text, figures, and supplementary datasets of the

444 original source material. In addition, we compiled datasets from unpublished sources using a
445 snowball search method, where we reached out to the authors of previously published kelp
446 papers in Canada and asked for recommendations on potential data sources for kelp extent,
447 biomass, canopy cover, and net primary productivity.

448
449 We focused subsequent analyses on the kelp species that had at least one biomass and NPP
450 record for a species on a given coast (Table C2). These included the two surface kelp species
451 (i.e., *Macrocystis pyrifera* and *Nereocystis luetkeana*) and seven of the 15 subsurface kelps
452 found on the Pacific coast (i.e., *Agarum clathratum*, *Costaria costata*, *Hedophyllum nigripes*,
453 *Neoagarum fimbriatum*, *Pterygophora californica*, *Pleurophyucus gardneri*, *Saccharina latissima*);
454 five of the seven species found on the Arctic coast (i.e., *A. clathratum*, *Laminaria digitata*, *L.*
455 *solidungula*, *H. nigripes*, and *S. latissima*; all subsurface); and three of the five species found on
456 the Atlantic coast (i.e., *A. clathratum*, *L. digitata*, and *S. latissima*; all subsurface). Since *H.*
457 *nigripes* could not be differentiated from *L. digitata* in some of the Arctic and Atlantic data
458 records, the two species were grouped together in subsequent analyses. Likewise, we grouped
459 *A. clathratum* and *N. fimbriatum* records due to the difficulties with differentiating these two
460 species in the field.

461

462 Determining the potential extent of kelp forests

463 We produced high, mid, and low estimates of the potential areal extent of subsurface kelp
464 forest in Canada using available depth, substrate, and kelp percent cover data. As a
465 hypothetical maximum potential extent for subsurface kelps, we calculated the area of suitable

466 rocky seafloor across Canada, i.e., the areal extent of rocky seafloor between mean low water
467 and the 20 m depth contour in millions of hectares (Mha)⁵⁰. This is a conservative depth cutoff
468 since kelp forests occur deeper (50 m) in some areas⁶⁶. Depth estimates were based on the
469 General Bathymetric Chart of the Ocean data (<https://www.gebco.net>; GEBCO)—a gridded
470 global terrain model for the ocean and land at 15-arc-second resolution. The extent of rocky
471 seafloor was based on public spatial data repositories for the Pacific and Atlantic coasts^{67,68}.
472 Since there was limited information on the distribution of rocky seafloor for most of the
473 Atlantic coast, we used the fraction of rocky seafloor found on the Scotian shelf (30.7 %) as a
474 coarse proxy⁶⁷. For the Arctic, we used the fraction of rocky seafloor used by previous global
475 studies (20%)¹⁸. Finally, we masked extents in the Arctic by the areal coverage of perennial sea
476 ice occurring at the northern edges, using spatial data layers from BioOracle⁶⁹.

477
478 To constrain the upper and lower bound estimates for the potential extent of subsurface kelp
479 forests, we combined the maximum potential extent maps with existing field surveys of the
480 percent cover of kelp forests from the Pacific, Atlantic and Arctic coasts. We acquired quadrat
481 surveys of the percent cover of subsurface kelp species from active monitoring programs^{70,71}
482 and the peer-reviewed literature (Table C1). To calculate the upper bound extent of subsurface
483 kelp forests, we multiplied the maximum potential extent by the 75th percentile of observed
484 percent cover for all kelp species (regardless of the species composition) on each coast.
485 Additionally, we calculated the lower bound as the maximum potential area multiplied by the
486 25th percentile of observed kelp cover.

487

488 We also produced high, mid, and low bound estimates for the potential areal extent of surface
489 kelp forests, as a particular case solely found on the Pacific coast of Canada. First, we calculated
490 the maximum potential extent of surface kelps as the area of suitable rocky seafloor above 10m
491 water depth—the depth above which 90% of the observations for bull kelp and giant kelps
492 occur in British Columbia (Fig. C3). To produce mid and low bound estimates for surface kelps,
493 we used available coarse maps derived from existing aerial- and satellite-remote sensing
494 products. We determined the shoreline distribution of *M. pyrifera* and *N. luetkeana* according
495 to their presence in georeferenced oblique aerial imagery collected and analyzed by the British
496 Columbia Shore Zone program from 2004 to 2007^{72,73}. From this dataset, we calculated the
497 upper bound of potential surface kelp extent as the intersection between previous shoreline
498 detections (i.e., within 500 m) and the maximum area of suitable rocky seafloor adjacent to the
499 shoreline. Finally, we acquired global distribution maps of surface canopy kelps determined
500 from classified 20m resolution Sentinel-2 satellite imagery from 2015 to 2019⁷⁴, which we used
501 as the lower bound extent of surface canopy forming species. We validated both datasets
502 through expert comparison with Google Earth Imagery, removing obvious false positives found
503 in higher estuaries and along the intertidal zone, in ArcGIS Pro Version. 3.0.

504

505 Determining the per-area carbon stocks of kelp species

506 We quantified carbon standing stocks associated with kelp forests in Canada by compiling
507 available data on the area-specific biomass and plant density of kelp species from published
508 and unpublished sources (Table C1). Wet weight measurements (i.e., g WW m⁻²) for each
509 species and coast were converted to dry weight (g DW m⁻²) using species- and coast- specific

510 conversions from the peer-reviewed literature⁵⁵. Kelp densities (i.e., number m⁻²) for each
511 species and coast were also converted to dry weight using available average individual weight
512 measurements⁵⁵. We then used species- and coast- specific ratios to convert dry weight
513 measurements to the area-specific organic carbon content (g C m⁻²) of each sample⁵⁵. Finally,
514 we converted all measurements of organic carbon content to carbon standing stocks in units of
515 megagrams per hectare (Mg ha⁻¹) by species.

516

517 Determining the per-area annual carbon production rates of kelp species

518 We used available published and unpublished measurements of net primary productivity for all
519 kelp species found in Canadian waters (Table C1)^{55,65}. We also used published net primary
520 productivity records from locations with similar environmental conditions to Canadian waters
521 (i.e., within the range of mean ocean temperatures observed on the Pacific, Arctic, and/or
522 Atlantic coasts according to BioOracle data layers)⁶⁹. All wet weights were converted to dry
523 weight measurements and then all dry measurements were converted to area-specific rates of
524 net primary productivity (i.e., g C m⁻² yr⁻¹), using species- and coast- specific conversions from
525 the literature⁵⁵. We then converted all measurements of NPP to annual carbon production in
526 units of megagrams per hectare per year (Mg ha⁻¹ Yr⁻¹) by species.

527

528 Bayesian hierarchical models

529 We used Bayesian hierarchical models to evaluate the potential for natural variation in the per-
530 area carbon stocks (Mg C ha⁻¹) and carbon production rates (Mg C ha⁻¹ yr⁻¹) of different kelp
531 species in Canada. Bayesian hierarchical models are parameterized similarly to hierarchical

532 linear regression models using the Stan computational framework (<http://mc-stan.org/>), which
533 can be accessed via the 'brms' package of the R programming language (version 2022.12).⁷⁵ A
534 key advantage of Bayesian hierarchical models is the ability to generate a posterior distribution,
535 representing the central tendency (i.e., the posterior mean) and the probabilistic range of
536 uncertainty surrounding a parameter estimate. A credible confidence interval (CCI) can be
537 derived from this posterior distribution, indicating the range within which the true parameter
538 value is likely to fall. Additionally, the 'brms' package allows for the explicit consideration
539 measurements of standard deviation as an additional response term, allowing for adjusted CCIs
540 that reflect greater uncertainty where there is greater variability in kelp biomass and NPP.
541 Finally, a Bayesian approach allows for the use of informative priors based on observations
542 from different systems that further constrain the parameter estimates and CCIs. Additional
543 information about using the R package "brms" can be found in the literature and the
544 documentation^{76,77}. Scripts for model parameterization, selecting informative priors, and
545 evaluating model outputs, can also found in our Github repository
546 ([https://github.com/jenmchenry1/A-blueprint-for-national-assessments-of-blue-carbon-
547 capacity-of-kelp-forests-CA](https://github.com/jenmchenry1/A-blueprint-for-national-assessments-of-blue-carbon-capacity-of-kelp-forests-CA)).

548
549 The observed carbon standing stocks and production rates of eleven kelp species were modeled
550 as the response variables. To account for measurement uncertainty in the available
551 observations, we included the standard deviation measurements representing the site-level
552 variation as an additional response term. For each species, we built sets of competing models
553 that tested the effects of different combinations of predictors on species carbon stocks and

554 production rates (Table C5). We accounted for potential fixed effects of mean sea surface
555 temperatures derived BioOracle⁶⁹ and the oceanic context (i.e., Pacific, Arctic, and Atlantic),
556 and controlled for the sampling year and site identity as random effects. The models ran for
557 5000 iterations with 2500 warm-ups using three chains. Convergence was visually assessed by
558 examining the trace plots and further verifying all coefficients achieving an Rhat value of 1⁷⁸. To
559 determine which model best described each response variable, we used an approximation for
560 leave-one-out (LOO) cross-validation ('loo' package)^{79,80}. We evaluated the performance of the
561 final models through a series of posterior predictive checks where draws from the posterior
562 distribution of model parameters were compared to the observed data as a measure of model
563 goodness of fit (Fig. C7 – C8).

564
565 Models were trained with weakly informative 'priors', setting the scale of the prior distribution
566 to be larger than and consistent with the range of potential observed values in our collated
567 response datasets (Table C6) and the range of global synthesized primary productivity
568 measurements from macroalgal forests⁶⁵. To ensure that our choice of priors did not overly
569 constrain the resulting posterior predictions or inflate the uncertainty intervals, we conducted a
570 prior sensitivity analysis for the three most data rich species in our dataset (i.e., *M. pyrifera*, *N.*
571 *leutkeana*, and *S. latissima*) and used the best matched set of priors for the remaining species.

572
573 We present the final models for eleven kelp species, which were selected by the approximate
574 LOO cross-validation (Table C7). The final models were used to generate posterior mean
575 estimates of the potential carbon stocks and production rates associated with kelp species,

576 including the 90% credible confidence intervals around those estimates. Significant differences
577 among the posterior mean estimates were assessed through comparison of percent overlap
578 between credible confidence intervals.

579

580 Estimating the national carbon stocks, production, and sequestration capacity of kelp forests

581 We estimated the standing carbon stocks (Tg C) of current kelp forests in Canada as the
582 summed product of the kelp forest extent (E_{coast}) and the carbon stock potential of kelp forests
583 across Canada's three coastlines ($C_{\text{Stock}_{\text{coast}}}$) (Equation 1). As inputs to this calculation, we used
584 the posterior mean estimates of the carbon stocks of individual kelp species (described above).
585 To account for the fact that kelps often persist in multi-species assemblages and thus are not
586 likely persisting at their maximum biomass potential, we estimated the per-area carbon stock of
587 kelp forests per coast ($C_{\text{Stock}_{\text{coast}}}$) as the summation of the posterior mean estimates for each
588 kelp species ($C_{\text{Stock}_{\text{spp}}}$), weighted by the relative abundance of that kelp species (A_{spp}), on each
589 coast. We used the maximum, upper, and low kelp forest extent estimates as inputs to
590 determine the most likely maximum, upper bound, and lower bound carbon stock potential of
591 each coast.

592

593 Equation 1.

594 **Total Standing Carbon Stock of Kelp Forests ($C_{\text{Stock}_{\text{total}}}$)**

$$595 \quad C_{\text{Stock}_{\text{total}}} = \sum (E_{\text{coast}1} \times C_{\text{Stock}_{\text{coast}1}}) + (E_{\text{coast}2} \times C_{\text{Stock}_{\text{coast}2}}) + \dots + (E_{\text{coast}N} \times C_{\text{Stock}_{\text{coast}N}})$$

596 **Carbon Stock of Kelp Forests Per-area ($C_{\text{Stock}_{\text{coast}}}$)**

$$597 \quad C_{\text{Stock}_{\text{coast}}} = \sum (C_{\text{Stock}_{\text{spp}1}} \times A_{\text{spp}1}) + (C_{\text{Stock}_{\text{spp}2}} \times A_{\text{spp}2}) + \dots + (C_{\text{Stock}_{\text{spp}N}} \times A_{\text{spp}N})$$

598

599 Additionally, we estimated the total annual carbon production capacity of kelp forests (Tg C yr⁻¹)
600 ¹) of current kelp forests in Canada as the summed product of the kelp forest extent (E_{coast}) and
601 the carbon production rate of kelp forests across Canada's three coastlines ($C_{\text{Prod}_{\text{coast}}}$)
602 (Equation 2). To estimate the per-area carbon production rate of kelp forests per coast
603 ($C_{\text{Prod}_{\text{coast}}}$), we summed the posterior mean estimates for each kelp species ($C_{\text{Prod}_{\text{spp}}}$),
604 weighted the relative abundance of that kelp species (A_{spp}), on each coast. We calculated the
605 total carbon production capacity of kelp forests per coast in terms of the maximum, upper
606 bound, and lower bound extent estimates.

607

608 Equation 2:

609 Total Standing Carbon Production of Kelp Forests ($C_{\text{Prod}_{\text{total}}}$)

$$610 \quad C_{\text{Prod}_{\text{total}}} = \sum (E_{\text{coast}1} \times C_{\text{Prod}_{\text{coast}1}}) + (E_{\text{coast}2} \times C_{\text{Prod}_{\text{coast}2}}) + \dots + (E_{\text{coast}N} \times C_{\text{Prod}_{\text{coast}N}})$$

611 Carbon Production of Kelp Forests Per-area ($C_{\text{Prod}_{\text{coast}}}$)

$$612 \quad C_{\text{Prod}_{\text{coast}}} = \sum (C_{\text{Prod}_{\text{spp}1} \times A_{\text{spp}1}}) + (C_{\text{Prod}_{\text{spp}2} \times A_{\text{spp}2}}) + \dots + (C_{\text{Prod}_{\text{spp}N} \times A_{\text{spp}N}})$$

613

614 Finally, we estimated the total annual capacity (Tg C yr⁻¹) for the export of kelp-derived carbon
615 beyond the continental shelf break (i.e., the 200-m isobath), as an approximation for the total
616 carbon sequestration in the deep ocean resulting from Canada's kelp forests (Equation 3). To do
617 so, we acquired estimates of the fraction of kelp carbon detritus ($\text{Exp}_{\text{ecoregion}}$) that may be
618 exported to the open ocean before decomposing according to a global model of shelf to open
619 ocean exchange for all ecoregions falling within Canada's EEZ³⁸. We determined the total

620 annual export capacity of kelp forests in Canada as the summed product of the estimated kelp
621 forest extent (E_{coast}) and the annual carbon export rate of kelp forests across Canada's three
622 coastlines ($C\text{Flux}_{\text{coast}}$) (Equation 3). As an input, we estimated $C\text{Flux}_{\text{coast}}$ as the summation of the
623 fraction of modeled hydrological export per ecoregion ($\text{Exp}_{\text{ecoregion}}$) multiplied by the per-area
624 annual carbon production of kelp forests for a given coast ($C\text{Prod}_{\text{coast}}$; calculated above). We
625 calculated the total carbon export capacity of kelp forests per coast in terms of the maximum,
626 upper bound, and lower bound extent estimates.

627

628 Equation 3:

629 Total Annual Export of Kelp Carbon to the Deep Ocean ($C\text{Flux}_{\text{total}}$)

630
$$C\text{Flux}_{\text{total}} = \sum (E_{\text{coast}1} \times C\text{Flux}_{\text{coast}1}) + (E_{\text{coast}2} \times C\text{Flux}_{\text{coast}2}) + \dots + (E_{\text{coast}N} \times C\text{Flux}_{\text{coast}N})$$

631 Annual Export of Kelp Carbon to the Deep Ocean Per Ecoregion ($\text{Exp}_{\text{ocean}}$)

632
$$C\text{Flux}_{\text{coast}} = \sum (C\text{Prod}_{\text{coast}1} \times \text{Exp}_{\text{ecoregion}1}) + (C\text{Prod}_{\text{coast}2} \times \text{Exp}_{\text{ecoregion}2}) + \dots + (C\text{Prod}_{\text{coast}N} \times \text{Exp}_{\text{ecoregion}N})$$

633

634

635 **Tables:**

636 **Table 1.** Comparison of the estimated extents, carbon stocks, carbon production rates, and carbon sequestration capacity of kelp
 637 forests, seagrass beds and salt marshes in Canada. Parenthetical values represent the lower and upper estimate values reported by
 638 this study and in the literature. *carbon sequestration for kelp forests is calculated in terms of the potential export of kelp detrital
 639 carbon to deep ocean sinks; carbon sequestration for seagrasses and salt marshes is calculated in terms of the amount of carbon
 640 accumulation in sediments. ND signifies no data for a particular field; NA signifies there the field is not applicable for a given
 641 ecosystem.

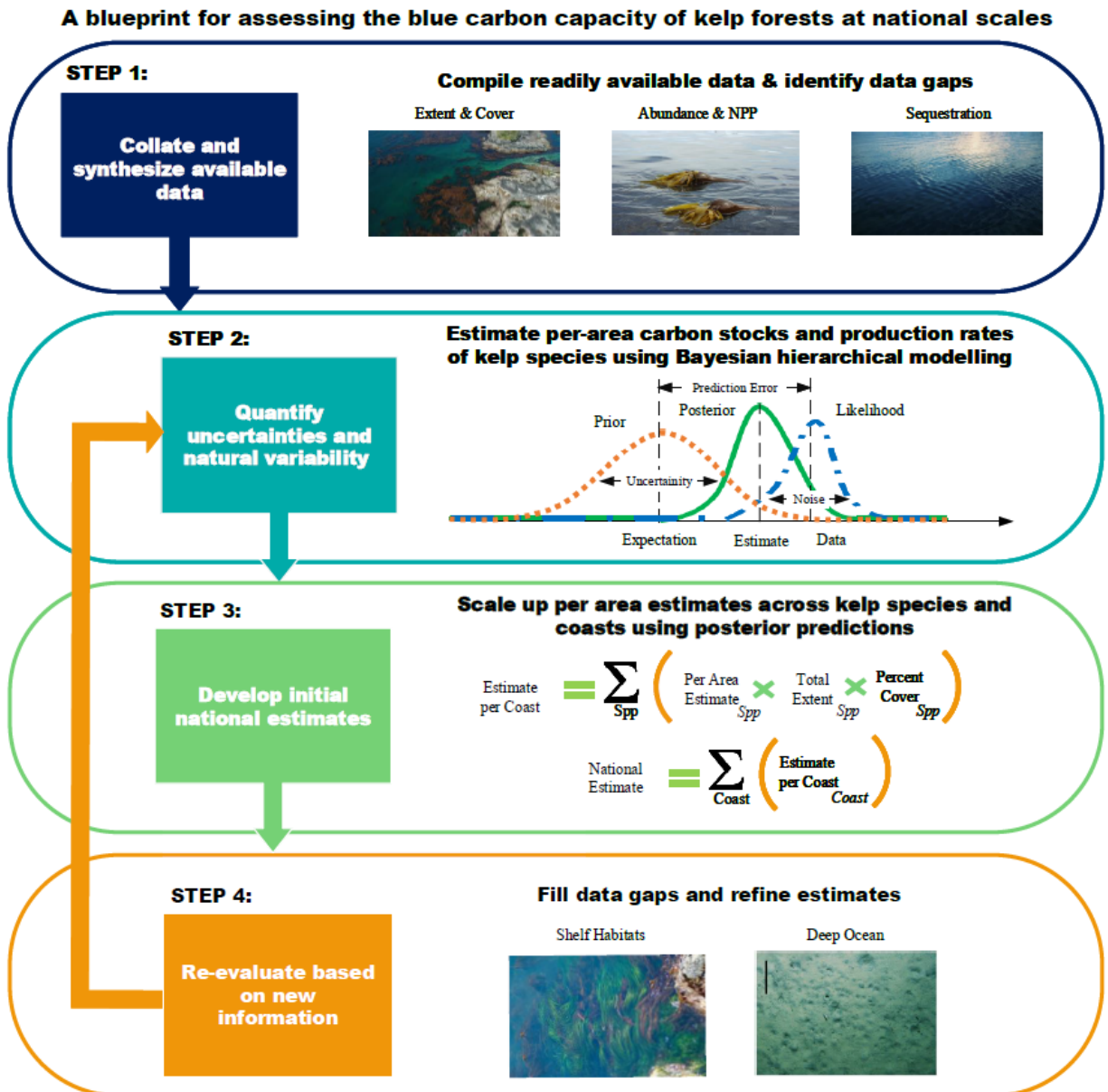
Ecosystem	Areal extent (Mha)	C stock per-area (Mg C ha⁻¹)	C production per-area (Mg C ha⁻¹ yr⁻¹)	C sequestration per-area * (Mg C ha⁻¹ yr⁻¹)	Standing C stock capacity (Tg C)	Standing C production capacity (Tg C yr⁻¹)	Standing C sequestration capacity* (Tg C yr⁻¹)
Kelp forests							
<i>Canopy:</i>	1.8 ⁺ (0.8 – 6.3)	0.8 ⁺ (0.4 – 1.2)	3.5 ⁺ (1.3 – 6.7)	0.6 ⁺ (0.3 – 1.5)	1.4 ⁺ (0.6 – 4.6)	3.1 ⁺ (1.1– 11.6)	0.2 ⁺ (0.1 – 1.2)
Seagrass meadows							
<i>Canopy:</i>	0.8 (0.2 – 1.4) ^a	0.1 (0.06 – 0.2) ^b	ND	NA	0.08 (0.01 – 0.3)	ND	NA
<i>Soils:</i>	0.8 (0.2 – 1.4) ^a	88.2 (50.2 – 380.1) ^c	NA	0.2 (0.04 – 0.9) ^{a,c}	70.6 (10.0 – 532.1)	NA	0.2 (0.01 – 1.3)
Salt marshes							
<i>Canopy:</i>	0.4 ^d	ND	ND	NA	ND	ND	NA
<i>Soils:</i>	0.4 ^d	80.4 (35.0 – 173) ^e	NA	2.0 (0.6 – 9.3) ^e	32.2	NA	0.8

Data sources: ⁺This study; ^aDrever et al. 2021, ^bPrentice et al. 2018, ^cPrentice et al. 2020, ^dRabinowitz & Andrews, ^eKelly et al. 2023.

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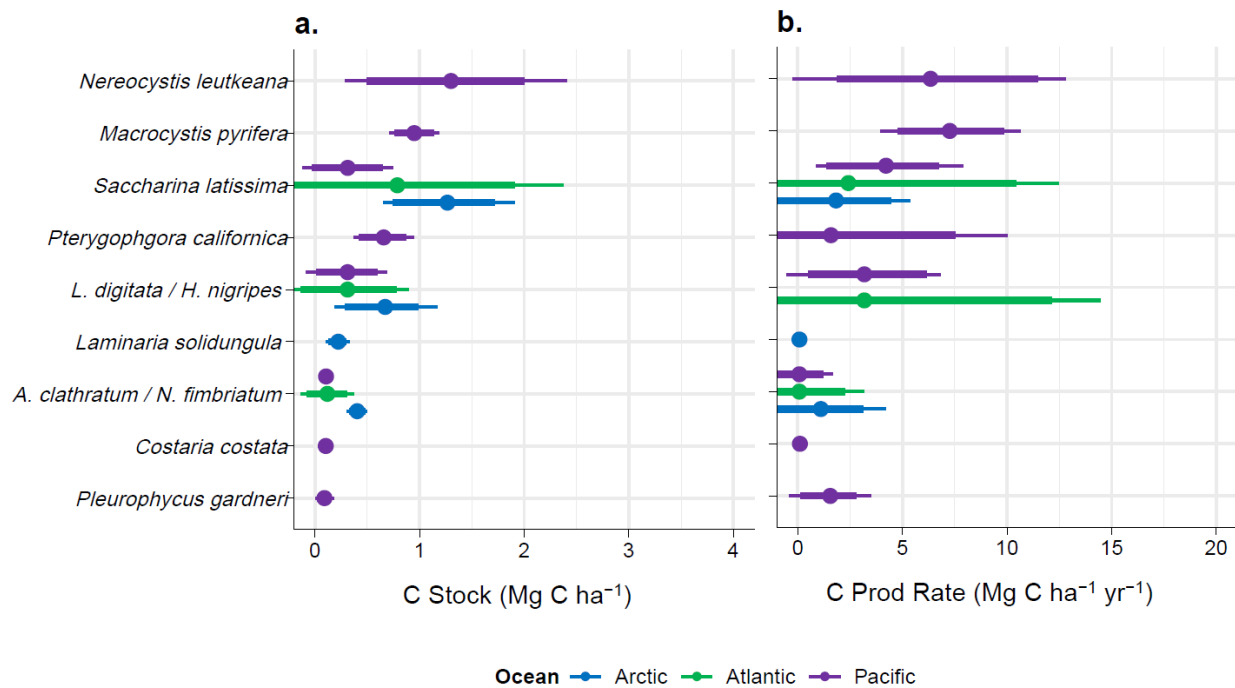
643 **Figures:**

644 **Figure 1.** A proposed blueprint for national assessments of the blue carbon capacity of kelp forests.
 645 Our proposed blueprint involves steps to 1) compile and synthesize available kelp data,
 646 2) quantify uncertainties and natural variability in potential rates of carbon production and
 647 storage by kelp species, 3) develop initial estimates of the carbon production, storage, and
 648 export capacity of kelp forests at national scales, and 4) refine assessments based on new
 649 information and data.



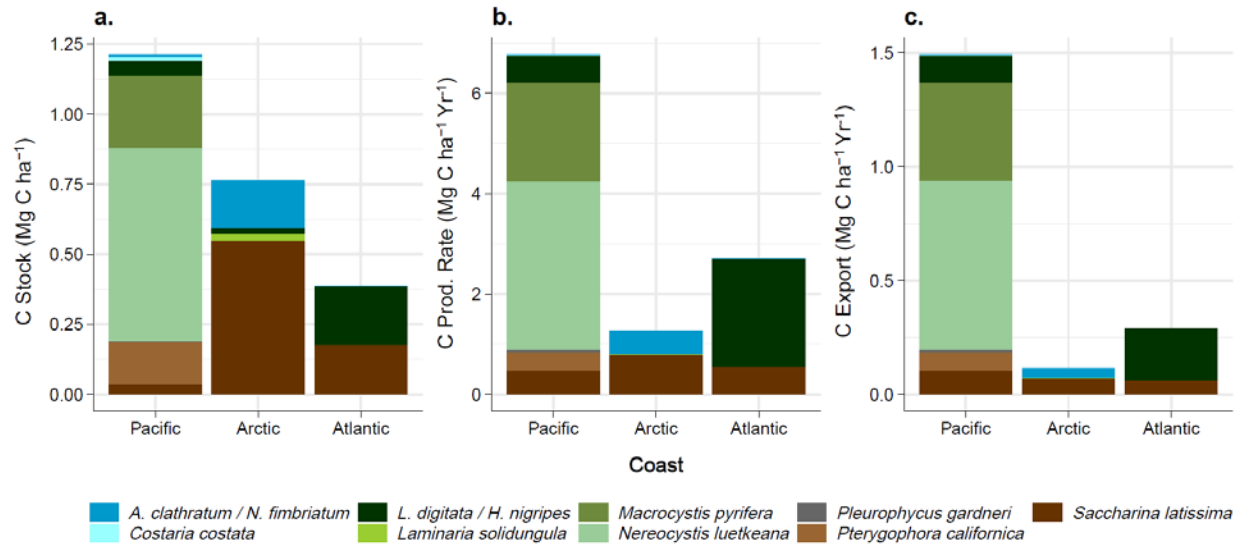
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653 **Figure 2.** Per- area estimates of the a) carbon stock (Mg C ha⁻¹) and b) carbon production (Mg C
 654 Ha⁻¹ yr⁻¹) capacity of kelp species across Canada's three coastlines (Pacific = purple; Arctic =
 655 blue; and Atlantic = green) according to Bayesian hierarchical models. Posterior mean estimates
 656 (and 90% credible intervals) are shown for each species, representing the average posterior
 657 predictive distribution conditional on the observed data and prior information. The inner and
 658 outer bars show the credible intervals representing the range of values within which the true
 659 mean estimates are likely to occur with 80% and 90% probability based on the final models.
 660 Kelp species include: *Macrocystis pyrifera*, *Nereocystis leutkeana*, *Costaria costata*, *Agarum*
 661 *clathratum* / *Neagarum fimbriatum*, *Laminaria digitata* / *Hedophyllum nigripes*, *Laminaria*
 662 *solidungula*, *Pterygophora californica*, *Pleurophycus gardneri*, and *Saccharina latissima*.



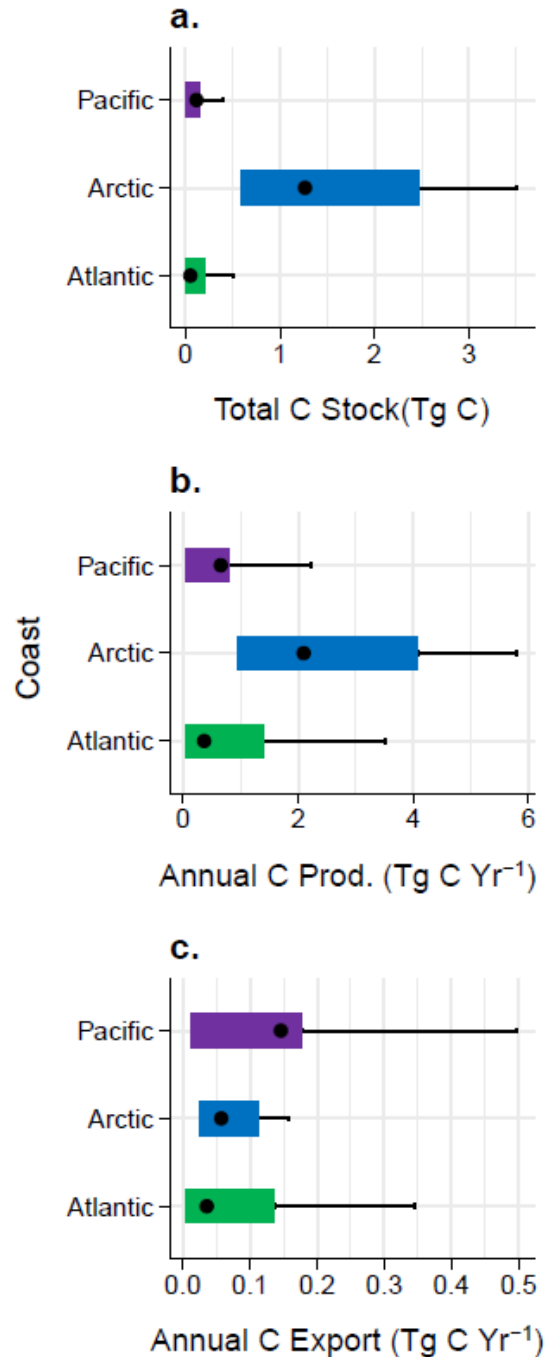
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668 **Figure 3.** Per-area posterior mean estimates of the a) carbon stocks (Mg C ha⁻¹), b) carbon
669 production (Mg C ha⁻¹ yr⁻¹), and c) carbon export (Mg C ha⁻¹ yr⁻¹) capacity of subtidal kelp
670 communities on Canada's three coasts. Stacked bar plots show the summed posterior means
671 across kelp species and coasts according to Bayesian hierarchical models, weighted by the
672 relative abundance of kelp species on each coast.
673



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682 **Figure 4.** National carbon capacity of Canadian kelp forests depicted in terms of the total
683 estimated a) standing carbon stocks (Tg C), b) carbon production (Tg C yr⁻¹), and c) carbon
684 export (Tg C yr⁻¹) capacity of kelp forests. The bars depict the upper bound (75th percentile) and
685 lower bound (25th percentile) estimates per coast. The circle represents the median estimates
686 for each coast. Error bars show the maximum potential capacity per coast.



687

688

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693 *IT28407*), The Tula Foundation and the Hakai Institute, ArcticNet (*ArcticKelp P101*) and the
694 Australian Research Council (*DP220100650*).

695

696 **Data Availability Statement:**

697 All research outputs, including the collated and synthesized datasets and the resulting national
698 estimates, will be made available at the time of publication through the data repository,
699 Borealis. Additionally, the R code supporting this study will be uploaded to a Github repository
700 ([https://github.com/jennmchenry1/A-blueprint-for-national-assessments-of-blue-carbon-](https://github.com/jennmchenry1/A-blueprint-for-national-assessments-of-blue-carbon-capacity-of-kelp-forests-CA)
701 [capacity-of-kelp-forests-CA](https://github.com/jennmchenry1/A-blueprint-for-national-assessments-of-blue-carbon-capacity-of-kelp-forests-CA)).

702

703 **References:**

- 704 1. Griscom, B. W. *et al.* Natural climate solutions. *Proc. Natl. Acad. Sci.* **114**, 11645–11650 (2017).
- 705 2. Macreadie, P. I. *et al.* Blue carbon as a natural climate solution. *Nat. Rev. Earth Environ.* **2**, 826–839
706 (2021).
- 707 3. Drever, C. R. *et al.* Natural climate solutions for Canada. *Sci. Adv.* **7**, eabd6034 (2021).
- 708 4. Fargione, J. E. *et al.* Natural climate solutions for the United States. *Sci. Adv.* **4**, eaat1869 (2018).
- 709 5. Alongi, D. M. Global Significance of Mangrove Blue Carbon in Climate Change Mitigation. *Sci* **2**, 67
710 (2020).
- 711 6. Chmura, G. L., Anisfeld, S. C., Cahoon, D. R. & Lynch, J. C. Global carbon sequestration in tidal, saline
712 wetland soils. *Glob. Biogeochem. Cycles* **17**, (2003).
- 713 7. Fourqurean, J. W. *et al.* Seagrass ecosystems as a globally significant carbon stock. *Nat. Geosci.* **5**,
714 505–509 (2012).
- 715 8. Mcleod, E. *et al.* A blueprint for blue carbon: toward an improved understanding of the role of
716 vegetated coastal habitats in sequestering CO₂. *Front. Ecol. Environ.* **9**, 552–560 (2011).
- 717 9. Bufarale, G. & Collins, L. B. Stratigraphic architecture and evolution of a barrier seagrass bank in the
718 mid-late Holocene, Shark Bay, Australia. *Mar. Geol.* **359**, 1–21 (2015).

- 719 10. Duarte, C. M., Middelburg, J. J. & Caraco, N. Major role of marine vegetation on the oceanic carbon
720 cycle. *Biogeosciences* **2**, 1–8 (2005).
- 721 11. Lo Iacono, C. *et al.* Very high-resolution seismo-acoustic imaging of seagrass meadows
722 (Mediterranean Sea): Implications for carbon sink estimates. *Geophys. Res. Lett.* **35**, (2008).
- 723 12. Waycott, M. *et al.* Accelerating loss of seagrasses across the globe threatens coastal ecosystems.
724 *Proc. Natl. Acad. Sci.* **106**, 12377–12381 (2009).
- 725 13. Macreadie, P. I. *et al.* Can we manage coastal ecosystems to sequester more blue carbon? *Front.*
726 *Ecol. Environ.* **15**, 206–213 (2017).
- 727 14. Pessarrodona, A. *et al.* Carbon sequestration and climate change mitigation using macroalgae: a
728 state of knowledge review. *Biol. Rev.* brv.12990 (2023) doi:10.1111/brv.12990.
- 729 15. Vanderklift, M. A. *et al.* A Guide to International Climate Mitigation Policy and Finance Frameworks
730 Relevant to the Protection and Restoration of Blue Carbon Ecosystems. *Front. Mar. Sci.* **9**, 872064
731 (2022).
- 732 16. Howard, J. *et al.* Blue carbon pathways for climate mitigation: Known, emerging and unlikely. *Mar.*
733 *Policy* **156**, 105788 (2023).
- 734 17. Krause-Jensen, D. & Duarte, C. M. Substantial role of macroalgae in marine carbon sequestration.
735 *Nat. Geosci.* **9**, 737–742 (2016).
- 736 18. Duarte, C. M. *et al.* Global estimates of the extent and production of macroalgal forests. *Glob. Ecol.*
737 *Biogeogr.* **31**, 1422–1439 (2022).
- 738 19. Jayatilake, D. R. M. & Costello, M. J. A modelled global distribution of the kelp biome. *Biol. Conserv.*
739 **252**, 108815 (2020).
- 740 20. Pedersen, M. F. *et al.* Detrital carbon production and export in high latitude kelp forests. *Oecologia*
741 **192**, 227–239 (2020).

- 742 21. Ortega, A. *et al.* Important contribution of macroalgae to oceanic carbon sequestration. *Nat. Geosci.*
743 **12**, 748–754 (2019).
- 744 22. Filbee-Dexter, K., Wernberg, T., Norderhaug, K. M., Ramirez-Llodra, E. & Pedersen, M. F. Movement
745 of pulsed resource subsidies from kelp forests to deep fjords. *Oecologia* **187**, 291–304 (2018).
- 746 23. Queirós, A. M. *et al.* Identifying and protecting macroalgae detritus sinks toward climate change
747 mitigation. *Ecol. Appl.* **33**, e2798 (2023).
- 748 24. Krumhansl, K. & Scheibling, R. Production and fate of kelp detritus. *Mar. Ecol. Prog. Ser.* **467**, 281–
749 302 (2012).
- 750 25. Hurd, C. L. *et al.* Forensic carbon accounting: Assessing the role of seaweeds for carbon
751 sequestration. *J. Phycol.* **58**, 347–363 (2022).
- 752 26. Paine, E. R., Schmid, M., Boyd, P. W., Diaz-Pulido, G. & Hurd, C. L. Rate and fate of dissolved organic
753 carbon release by seaweeds: a missing link in the coastal ocean carbon cycle. *J. Phycol.* **57**, 1375–
754 1391 (2021).
- 755 27. Geraldi, N. R. *et al.* Fingerprinting blue carbon: rationale and tools to determine the source of
756 organic carbon in marine depositional environments. *Front. Mar. Sci.* **263** (2019).
- 757 28. Krause-Jensen, D. *et al.* Sequestration of macroalgal carbon: the elephant in the Blue Carbon room.
758 *Biol. Lett.* **14**, 20180236 (2018).
- 759 29. Queirós, A. M. *et al.* Connected macroalgal-sediment systems: blue carbon and food webs in the
760 deep coastal ocean. *Ecol. Monogr.* **89**, (2019).
- 761 30. Fujita, R. *et al.* Seaweed blue carbon: Ready? Or Not? *Mar. Policy* **155**, 105747 (2023).
- 762 31. Araya-Lopez, R., de Paula Costa, M. D., Wartman, M. & Macreadie, P. I. Trends in the application of
763 remote sensing in blue carbon science. *Ecol. Evol.* **13**, e10559 (2023).
- 764 32. Mora-Soto, A. *et al.* Environmental characteristics and floating kelp dynamics in the southern Salish
765 Sea, British Columbia, Canada. *Front. Mar. Sci.* (in review).

- 766 33. Archambault, P. *et al.* From sea to sea: Canada’s three oceans of biodiversity. *PLoS One* **5**, e12182
767 (2010).
- 768 34. Ross, F. W. *et al.* A preliminary estimate of the contribution of coastal blue carbon to climate change
769 mitigation in New Zealand. *N. Z. J. Mar. Freshw. Res.* 1–11 (2023).
- 770 35. Filbee-Dexter, K. & Scheibling, R. E. Sea urchin barrens as alternative stable states of collapsed kelp
771 ecosystems. *Mar. Ecol. Prog. Ser.* **495**, 1–25 (2014).
- 772 36. Thompson, M. MaPP Kelp Monitoring Protocol. (2021).
- 773 37. Filbee-Dexter, K. *et al.* Ocean temperature controls kelp decomposition and carbon sink potential.
774 *PLOS Biol.* (2020).
- 775 38. Filbee-Dexter, K. *et al.* Carbon export from seaweed forests to deep ocean sinks. (in review).
- 776 39. Connell, S. D. *et al.* Recovering a lost baseline: missing kelp forests from a metropolitan coast. *Mar.*
777 *Ecol. Prog. Ser.* **360**, 63–72 (2008).
- 778 40. Filbee-Dexter, K., Feehan, C. & Scheibling, R. Large-scale degradation of a kelp ecosystem in an
779 ocean warming hotspot. *Mar. Ecol. Prog. Ser.* **543**, 141–152 (2016).
- 780 41. Johnson, C. R. *et al.* Climate change cascades: Shifts in oceanography, species’ ranges and subtidal
781 marine community dynamics in eastern Tasmania. *J. Exp. Mar. Biol. Ecol.* **400**, 17–32 (2011).
- 782 42. Steneck, R. S. *et al.* Kelp forest ecosystems: biodiversity, stability, resilience and future. *Environ.*
783 *Conserv.* **29**, 436–459 (2002).
- 784 43. Wernberg, T. *et al.* An extreme climatic event alters marine ecosystem structure in a global
785 biodiversity hotspot. *Nat. Clim. Change* **3**, 78–82 (2013).
- 786 44. Starko, S. *et al.* Microclimate predicts kelp forest extinction in the face of direct and indirect marine
787 heatwave effects. *Ecol. Appl.* **32**, (2022).
- 788 45. Steneck, R. S., Leland, A., McNaught, D. C. & Vavrinec, J. Ecosystem flips, locks, and feedbacks: the
789 lasting effects of fisheries on Maine’s kelp forest ecosystem. *Bull. Mar. Sci.* **89**, 31–55 (2013).

- 790 46. Attridge, C. M., Metaxas, A. & Denley, D. Wave exposure affects the persistence of kelp beds amidst
791 outbreaks of the invasive bryozoan *Membranipora membranacea*. *Mar. Ecol. Prog. Ser.* **702**, 39–56
792 (2022).
- 793 47. Oliver, E. C. *et al.* Projected marine heatwaves in the 21st century and the potential for ecological
794 impact. *Front. Mar. Sci.* **6**, 734 (2019).
- 795 48. Assis, J., Araújo, M. B. & Serrão, E. A. Projected climate changes threaten ancient refugia of kelp
796 forests in the North Atlantic. *Glob. Change Biol.* **24**, e55–e66 (2018).
- 797 49. Jueterbock, A. *et al.* Climate change impact on seaweed meadow distribution in the North Atlantic
798 rocky intertidal. *Ecol. Evol.* **3**, 1356–1373 (2013).
- 799 50. Filbee-Dexter, K. & Wernberg, T. Substantial blue carbon in overlooked Australian kelp forests. *Sci.*
800 *Rep.* **10**, 12341 (2020).
- 801 51. Goldsmit, J. *et al.* Kelp in the Eastern Canadian Arctic: Current and Future Predictions of Habitat
802 Suitability and Cover. *Front. Mar. Sci.* **18**, 742209 (2021).
- 803 52. Wright, L., Pessarrodona, A. & Foggo, A. Climate-driven shifts in kelp forest composition reduce
804 carbon sequestration potential. *Glob. Change Biol.* (2022).
- 805 53. Filbee-Dexter, K., Wernberg, T., Fredriksen, S., Norderhaug, K. M. & Pedersen, M. F. Arctic kelp
806 forests: Diversity, resilience and future. *Glob. Planet. Change* **172**, 1–14 (2019).
- 807 54. Starko, S. *et al.* Temperature and food chain length, but not latitude, explain region-specific kelp
808 forest responses to an unprecedented heatwave. *bioRxiv* 2023–01 (2023).
- 809 55. Pessarrodona, A., Filbee-Dexter, K., Krumhansl, K. A., Moore, P. J. & Wernberg, T. A global dataset of
810 seaweed net primary productivity. *Sci. Data* (2022) doi:10.1101/2021.07.12.452112.
- 811 56. Filbee-Dexter, K. *et al.* *in review*. Seaweed forests are carbon sinks that can mitigate CO2 emissions.
812 Nature Geosciences.

- 813 57. Leithold, E. L., Blair, N. E. & Wegmann, K. W. Source-to-sink sedimentary systems and global carbon
814 burial: A river runs through it. *Earth-Sci. Rev.* **153**, 30–42 (2016).
- 815 58. Hurd, C., Gattuso, J. & Boyd, P. W. Air-sea carbon dioxide equilibrium: Will it be possible to use
816 seaweeds for carbon removal offsets? *J. Phycol.* (2023).
- 817 59. Needelman, B. A. *et al.* The science and policy of the verified carbon standard methodology for tidal
818 wetland and seagrass restoration. *Estuaries Coasts* **41**, 2159–2171 (2018).
- 819 60. Starko, S. *et al.* A comprehensive kelp phylogeny sheds light on the evolution of an ecosystem. *Mol.*
820 *Phylogenet. Evol.* **136**, 138–150 (2019).
- 821 61. Druehl, L. D. The pattern of Laminariales distribution in the northeast Pacific. *Phycologia* **9**, 237–247
822 (1970).
- 823 62. Bringloe, T. T., Verbruggen, H. & Saunders, G. W. Population structure in Arctic marine forests is
824 shaped by diverse recolonisation pathways and far northern glacial refugia. *bioRxiv* 2020–03 (2020).
- 825 63. Bolton, J. J. The biogeography of kelps (Laminariales, Phaeophyceae): a global analysis with new
826 insights from recent advances in molecular phylogenetics. *Helgol. Mar. Res.* **64**, 263–279 (2010).
- 827 64. Assis, J. *et al.* A fine-tuned global distribution dataset of marine forests. *Sci. Data* **7**, 119 (2020).
- 828 65. Pessarrodona, A. *et al.* Global seaweed productivity. *Sci. Adv.* **8**, eabn2465 (2022).
- 829 66. Castro de la Guardia, L. *et al.* Increasing depth distribution of Arctic kelp with increasing number of
830 open water days with light. *Elem Sci Anth* **11**, 00051 (2023).
- 831 67. Greenlaw, M. & Harvey, C. Data of: A substrate classification for the Inshore Scotian Shelf and Bay of
832 Fundy, Maritimes Region. (2022).
- 833 68. Gregr, E. J., Haggarty, D. R., Davies, S. C., Fields, C. & Lessard, J. Comprehensive marine substrate
834 classification applied to Canada’s Pacific shelf. *PLOS ONE* **16**, e0259156 (2021).
- 835 69. Assis, J. *et al.* Bio-ORACLE v2. 0: Extending marine data layers for bioclimatic modelling. *Glob. Ecol.*
836 *Biogeogr.* **27**, (2018).

- 837 70. Dive Survey Algae And Substrate Data. OpenData Fisheries & Oceans Canada (2024).
- 838 71. Krumhansel, K. Unpublished data. (2022).
- 839 72. British Columbia Shore Zone Data Portal.
- 840 <https://soggy2.zoology.ubc.ca/geonetwork/srv/eng/catalog.search#/metadata/501b77d2-573c->
- 841 [44d9-ac92-140d0f024e44](https://soggy2.zoology.ubc.ca/geonetwork/srv/eng/catalog.search#/metadata/501b77d2-573c-44d9-ac92-140d0f024e44) (2022).
- 842 73. Howes, D., Harper, J. & Owens, E. Physical shore-zone mapping system for British Columbia. *Rep.*
- 843 *Prep. Environ. Emerg. Serv. Minist. Environ. Vic. BC Coast. Ocean Resour. Inc* Sidney BC Owens Coast.
- 844 *Consult. Bainbridge WA* (1994).
- 845 74. Mora-Soto, A. *et al.* A high-resolution global map of giant kelp (*Macrocystis pyrifera*) forests and
- 846 intertidal green algae (*Ulvophyceae*) with Sentinel-2 imagery. *Remote Sens.* **12**, 694 (2020).
- 847 75. Bürkner, P.-C. Advanced Bayesian multilevel modeling with the R package brms. *ArXiv Prepr.*
- 848 *ArXiv170511123* (2017).
- 849 76. Bürkner, P.-C. brms: An R package for Bayesian multilevel models using Stan. *J. Stat. Softw.* **80**, 1–28
- 850 (2017).
- 851 77. Barreda, S. & Silbert, N. Fitting Bayesian regression models with brms. in *Bayesian Multilevel Models*
- 852 *for Repeated Measures Data* 55–86 (Routledge, 2023).
- 853 78. Gelman, A. Parameterization and Bayesian modeling. *J. Am. Stat. Assoc.* **99**, 537–545 (2004).
- 854 79. Vehtari, A., Gelman, A. & Gabry, J. Practical Bayesian model evaluation using leave-one-out cross-
- 855 validation and WAIC. *Stat. Comput.* **27**, 1413–1432 (2017).
- 856 80. Vehtari, A. *et al.* loo: Efficient leave-one-out cross-validation and WAIC for Bayesian models. (2023).
- 857